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دراسة عملية للتخلص من الملون الكاثيوني (الأحمر القاعدي 12) من المحاليل المائية عن طريق الامتصاص بواسطة بقايا العظام الحيوانية

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المخلص:

تطُرقت الدراسة إلى فاعلية استخدام بقايا العظام الحيوانية المتواجدة بوفرة في بيئتنا، كمادة ماصة، منخفضة التكلفة لتحضيرها للتخلص من الملون الكاثيوني (الأحمر القاعدي 12) من المحاليل المائية وخصوصا المستعملة في قطاع النسيج. لقد قمنا بدراسة مدى تأثير بعض العوامل التجريبية، كمية المادة الماصة، زمن العملية، تركيز الملون، قابلية الحموضة ودرجة الحرارة على قابلية التخلص من الملون الأحمر. لقد أظهرت نتائج الدراسة بأنه قد تم التخلص من المادة الملونة بفاعلية بلغت 83.56% وذلك باستعمال 50 ملغ من مادة بقايا العظام وكان الاتزان بين المادتين الماصة والملونة سريع وتم الوصول إليه بعد زمن يقارب 60 دقيقة. لقد تم تحليل بيانات الإمتصاص الحراري باستخدام نماذج لنكمير وفرونديك الخاصة بالإمتصاص الحراري، وكانت قدرة الامتصاص Q_m حسب طريقة لنكمير قد وصلت إلى 64.13، 67.87، 66.73 ملغ/غ عند درجات الحرارة التالية، 303، 313، 323، 333 كيلفن (°K). من خلال جميع المعطيات الفيزيوكيميائية تبين أن المادة المستخلصة من بقايا العظام الحيوانية، يمكن أن تستخدم كمادة ماصة، منخفضة التكلفة للتخلص من المواد الملونة والمستعملة أساسا في ميدان النسيج.



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ORIGINAL ARTICLE

Removal of a cationic dye – Basic Red 12 – from aqueous solution by adsorption onto animal bone meal

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KEYWORDS

Animal bone meal;
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Thermodynamic studies

Abstract The efficiency of animal bone meal as a low cost adsorbent for removing a cationic dye Basic Red 12 from aqueous solution has been investigated. The effects of various experimental factors; adsorbent dose, contact time, dye concentration, pH and temperature were studied by using the batch technique. A maximum of 83.56% removal of the dye was observed by 50 mg of Animal Bone Meal (ABM). Dye adsorption equilibrium was rapidly attained after 60 min of contact time. The isotherms of adsorption data were analyzed by Langmuir and Freundlich adsorption isotherm models. The adsorption capacity, Q_m , obtained from the Langmuir isotherm plots were 64.13, 64.95, 66.73 and 67.87 mg/g respectively at 303, 313, 323 and 333 K. The characteristic results; dimensionless separation factor, R_L , and the adsorption intensity, n , showed that animal bone meal can be employed as an alternative to commercial adsorbents in the removal of Basic Red 12 from aqueous solution.

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1. Introduction

Pollution from chemicals products is a real threat to the aquatic environment, air, and soil. Several investigations were conducted and tireless efforts were made to fight against pollution in general. In this context, we have contributed to this research field especially the water pollution by textile dyes. In this fact, the discharge of colored wastewater from industry into natural stream causes many significant problems such as increasing the toxicity (Pearce et al., 2003; Ravi et al., 2005) and chemical oxygen demand of effluent as well as reducing light penetration (Yasemin and Haluk, 2006). It is recognized that public perception of water quality is greatly influenced by the color. Color is the first contaminant to be recognized in wastewater (Banat et al., 1996). The presence of very small amounts of

dyes in water less than one ppm for some dyes is highly visible and undesirable (Banat et al., 1996; Robinson et al., 2001).

Several treatment processes are in practice for the removal of dyes from wastewater. These processes include; solar photo-Fenton degradation, photo catalytic degradation, photo-Fenton processes, biodegradation, integrated chemical-biological degradation, electrochemical degradation, coagulation and flocculation, oxidation and ozonation, membrane separation, sonochemical degradation and adsorption.

Among treatment technologies, adsorption is rapidly gaining prominence. Activated carbon is the most widely used adsorbent for dye removal, but it is too expensive (Malik, 2003), consequently, numerous low-cost alternative adsorbents have been proposed including: chemically modified sugarcane bagasse lignin (Da Silva et al., 2011), Pistachio hull waste (Moussavi and Khosravi, 2011), Coffee husk-based activated carbon (Ahmad and Rahman, 2011), Pine cone (Mahmoodi et al., 2011), Rice husk (Safa and Bhatti, 2011), Synthetic calcium phosphates (Boujaady et al., 2011), Natural untreated clay (Errais et al., 2011), Pillared clays (Gil et al., 2011), Swelling clays (Li et al., 2011).

This work describes the feasibility of removal of a cationic dye – Basic Red 12 – from aqueous solutions by using animal bone meal as a new low cost adsorbent. Batch adsorption experiments were achieved out as a function of adsorbent dosage, contact time, initial concentration of dye, pH and temperature. The thermodynamic parameters were also evaluated from the adsorption measurements. The Langmuir and Freundlich isotherms models, kinetics adsorption are also studied.

2. Materials and methods

2.1. Adsorbate

Basic Red 12 (BR12) was the cationic dye used in this study, supplied by SDI textile company (Safi, Morocco) and was used without any purification. The chemical structure of BR12 is given in Fig. 1. Colored solutions were prepared by dissolving required quantity of BR12 in distilled water. The final volume prepared was 500 mL. Adsorption studies for the evaluation of ABM adsorbent for the removal of BR12 dye from aqueous solutions were carried out in triplicate to get concordant results using a batch contact adsorption method.

2.2. Preparation of adsorbent

Animal bones were collected from nearby butcher shops. All of the attached meat and fat were removed and cleaned from the bones. The bones were then washed several times with tap

water and left in open air for several days to get rid of odors. Later, they were transferred to the oven at 80 °C for drying. The dried bones were crushed and milled into different particle sizes in the range 45–200 µm then calcined for 2 h at 800 °C. The residue was washed with water and was used after drying for 24 h at 80 °C. The residue was washed with water and was dried overnight at 100 °C in a conventional drying oven, and then calcined at a heating rate of 2 °C/min to 400 °C and kept at this temperature for 4 h. The calcined material was ground to a fine powder in an electrical grinder. This was then passed through a mesh screen to get particles < 500 µm size. This was stored in air tight container for further use.

2.3. Determination of pH zero charge point

The zero point charge pH (pH_{ZPC}) of the ABM adsorbent was measured using the pH drift method (Calvete et al., 2009). In this fact, the pH_{ZPC} of the adsorbent was determined by adding 20 mL of 5×10^{-2} mol/L NaCl to several 50 mL cylindrical high-density polystyrene flasks (height 117 mm and diameter 30 mm). A range of initial pH (pH_i) values of the NaCl solutions were adjusted from 2 to 12 by adding 10^{-1} mol/L of HCl and NaOH. The total volume of the solution in each flask was brought to exactly 30 mL by further addition of 5×10^{-2} mol/L NaCl solution. The pH_i values of the solutions were then accurately noted and 50 mg of each adsorbent were added to each flask, which was securely capped immediately. The suspensions were shaken in a shaker at 298 K and allowed to equilibrate for two days. The suspensions were then centrifuged at 3600 rpm for 15 min and the final pH (pH_f) values of the supernatant liquid were recorded. The value of pH_{ZPC} is the point where the curve of ΔpH ($pH_f - pH_i$) versus pH_i crosses the line equal to zero. Measured pH was down by pH-Metre Basic 20+ model pH-meter.

2.4. Experimental procedure

The ABM adsorbent was characterized using elemental analysis, FTIR and XRD. FTIR spectra were obtained using ATI Mattson Genesis series FTIR™ UNICAM instrument. XRD of ABM adsorbent was carried out by using X pert Pro X-ray diffractometer using copper anticathode $\lambda(Cu) = 1.5418 \text{ \AA}$.

The surface area of ABM was determined by BET method and the calculated surface area was 85 m²/g. The concentration of BR12 before and after adsorption was determined using UV-Visible spectrophotometer (Biomate 6, England) at $\lambda_{max} = 469 \text{ nm}$.

Preliminary experiments demonstrated that the equilibrium was established at 60 min. A 50 mg sample of ABM was mixed with 50 mL dye solution of appropriate concentration (20, 40, 60, 80 and 100 mg/L). Samples of 10 mL of mixture were withdrawn from the batch at predetermined time intervals and the supernatant was centrifuged for 15 min at 3600 rpm. All dye solutions prepared were filtered by Millipore membrane type 0.45 µm HA. An incubator shaker was used for carrying out experiments at 303, 313, 323 and 333 K.

The amount of adsorption at time t , Q_t (mg/g) was calculated using the following formula:

$$Q_t = \frac{C_0 - C_t}{W} V \quad (1)$$

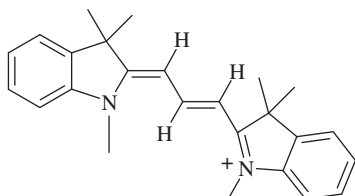


Figure 1 Chemical structure of Basic Red 12.

where C_t (mg/L) is the liquid concentration of dye at any time, C_0 (mg/L) is the initial concentration of the dye in solution, V is the volume of the solution (L) and W is the mass of dye adsorbent (g).

The amount of equilibrium adsorption Q_e (mg/g) was calculated using the formula:

$$Q_e = \frac{C_0 - C_e}{W} V \quad (2)$$

where C_0 and C_e (mg/L) are the liquid concentrations of dye initially and at equilibrium.

The dye removal percentage can be calculated as follows:

$$\% \text{ of dye removal} = \frac{C_0 - C_e}{C_0} \times 100 \quad (3)$$

where C_0 and C_e (mg/L) are the initial and equilibrium concentrations of dye in solution.

3. Results and discussion

3.1. Characterization of ABM adsorbent

Elemental analysis of ABM shows a high yield of Ca (49.62%) and P (42.36%) with a (Ca/P) ratio equal to 1.17. Small amounts of Si (3.88%), Mg (1.32%), Na (0.77%), Al (0.35%), Fe (0.24%), Cl (0.24%), S (0.11%), K (0.07%), Sr (0.03%), Cu (0.03%) and Zn (0.02%) are found. Fig. 2 shows the FTIR spectra of ABM adsorbent. FTIR spectra was carried out by encapsulating 0.5 mg of finely powder with 400 mg of KBr to get translucent disk. Carbonate is the most abundant substitution in bone mineral and according to its crystal position, carbonate apatite is designated as type A (OH^-) or type B (PO_4^{3-}), the latter being the most frequent bone (Best et al., 2008; Landi et al., 2003). It is recognized that hydroxyapatite derived from natural bone (Tadic and Epple, 2004; Figueiredo et al., 2009). FTIR spectrum of ABM given in Fig. 2 shows the characteristic bands of hydroxyapatite

(571, 603, 962 (shoulder), 1047 and 1091 (shoulder) cm^{-1} due to phosphate vibrations and collagen ($\text{C}=\text{O}$ stretching vibration at 1635 cm^{-1} , N-H in plane bending at 1458 cm^{-1} , C-H and N-H stretching modes in $3000\text{--}3571 \text{ cm}^{-1}$ region) (Mkukuma et al., 2004; Paschalis et al., 1997; Rehman and Bonfield, 1997). Additionally, the typical bands of carbonate substituting for phosphate site (type B) in the apatite lattice are also observed: band at 874 cm^{-1} and double bands $1385/1445 \text{ cm}^{-1}$ (Murugan et al., 2006; Haberko et al., 2006). Fig. 3 depicts the XRD spectrum of ABM adsorbent. The diffractogram of calcined ABM shows only the characteristic pattern of hydroxyapatite. These results confirm that the amorphous organic component was removed after calcination as found in the literature (Figueiredo et al., 2010). For 2θ between 25° and 45° , the main lattice reflections originate peaks at 25 , 28.1 , 32.8 , 33.7 , 34.5 and 39.7° being respectively assigned to the (002), (102), (210), (211), (112) and (300) Miller plans of hydroxyapatite (Figueiredo et al., 2010).

3.2. Effect of adsorbent amount

To evaluate the effect of adsorbent dose on the adsorption of BR12, we have carried out equilibrium experiments by contacting different amounts of ABM from the range of (10–90 mg) with 50 mL of a solution dye at 60 mg/L as initial concentration at 30°C . The agitation was made for 24 h. Fig. 4 depicts the variation of percentage of removal BR12 versus adsorbent dose. The percentage of removal BR12 was increased with increasing adsorbent dose. A maximum of 83.56% removal of the dye was observed by 50 mg of ABM.

3.3. Effect of contact time and DR12 dye concentration

In order to achieve the effect of contact time and initial BR12 concentration (20–100 mg/L) with 50 mg of ABM on the percentage of removal BR12 by ABM, different experiments were realized. Fig. 5 depicts these results. It was observed that the re-

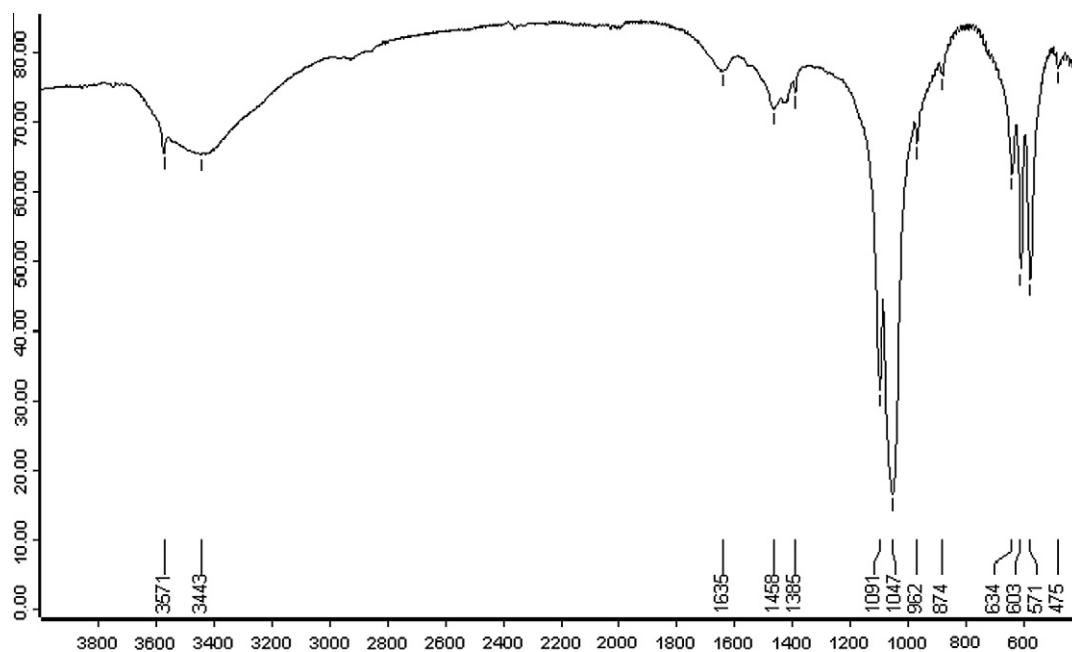


Figure 2 IR spectra of ABM.

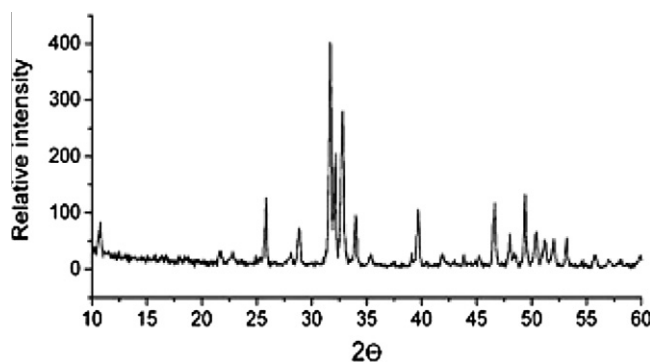


Figure 3 X-ray diffraction of ABM.

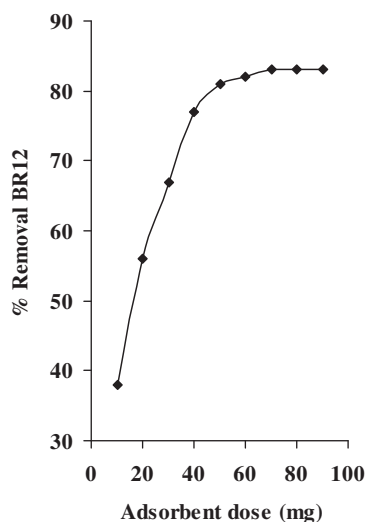


Figure 4 Effect of adsorbent dose on the adsorption of BR12: [BR12] = 60 mg/L; contact time = 24 h.

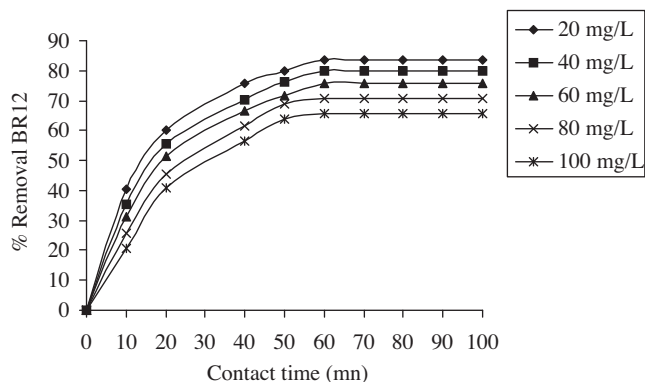


Figure 5 Effect of contact time and dye concentration adsorbent dose: $W = 50$ mg and $V = 50$ mL.

removal of BR12 is increased with increasing the contact time at all initial BR12 dye concentrations. Furthermore, the % removal dye is increased with the decrease in initial dye concentration. It is because of the fact at lower concentration, the ratio of the initial number of dye molecules to the available surface area is low subsequently the fractional adsorption becomes indepen-

dent of initial concentration. However, at high concentration the available sites of adsorption becomes fewer and hence the percentage removal of dye depends upon concentration. For the first 30 min, the adsorption uptake was rapid then it proceeds at a slower adsorption rate and finally it attains saturation at 60 min. The obtained removal curves were single, smooth and continuous, indicating monolayer coverage of dye on the surface of adsorbent (Al-Rashed and Al-Gaid, 2012).

3.4. Effect of pH

One of the most important factors in adsorption studies is the effect of acidity on the medium (Calvete et al., 2010). Different species may present divergent ranges of suitable pH depending on which adsorbent is used. The effect of initial pH on the adsorption capacity of BR12 dye using ABM adsorbent was evaluated within pH range between 2 and 12. The mean of the duplicated experimental results is plotted in Fig. 6, indicating that pH considerably affected BR12 adsorption, particularly under acidic conditions. As can be seen in Fig. 6, percentage removal of BR12 onto ABM increased from 24.67% to 83.23% when the solution pH was increased from 2 to 12. The reason that ABM behaved differently in adsorbing BR12 at different solution pHs can be explained by considering the pH_{ZPC} as shown in Fig. 7 of the adsorbent as well as molecular nature of BR12 (cationic dye). The pH_{ZPC} of ABM particles is 8.4, meaning that the adsorbent's surface was positively charged at solution pH below 8.4. This causes competition between protons and BR12 formed cations for adsorption locations as well as the repulsion of cationic BR12 molecules, resulting in the reduction of dye adsorption. The lower the pH goes below pH_{ZPC} , the greater the density of positive ions on the surface of ABM will be which in turn allows for less adsorption. This is confirmed by low BR12 removal at a strong acidic pH of 2. When solution pH increases above pH_{ZPC} , a negative charge is present on the surface of ABM, causing better BR12 cations adsorption through the electrostatic attraction phenomenon. The maximum of BR12 adsorption onto ABM was achieved above pH 9.1.

3.5. Adsorption isotherms and thermodynamic studies

The equilibrium existence of dye between the liquid phase and the solid phase is well described by adsorption isotherms. Experimental isotherm data collected at different temperatures

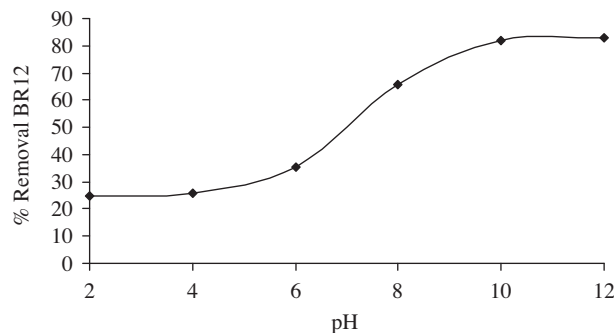


Figure 6 Effect of pH on removal of BR12 onto ABM. Initial dye concentration 60 mg/L. Adsorbent dose 50 mg, volume solution 50 mL, agitation time 24 h, temperature 25 °C.

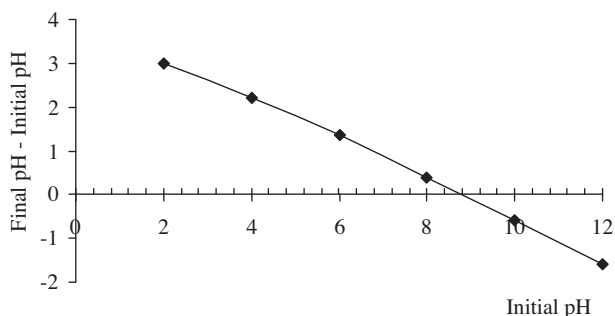


Figure 7 Determination of zero point charge pH.

303, 313, 323 and 333 K fit in Langmuir and Freundlich adsorption isotherm models and thermodynamic parameters calculated accordingly. For the equilibrium concentration of adsorbate C_e and amount of dye adsorbed at equilibrium Q_e , the following linear forms of Langmuir (Langmuir, 1916) and Freundlich (Butt and Graf, 2003) isotherms were studied.

$$\frac{C_e}{Q_e} = \frac{C_e}{Q_m} + \frac{1}{Q_m} b \quad (4)$$

$$\log Q_e = \log K_f + \frac{1}{n \log C_e} \quad (5)$$

where C_e is the equilibrium concentration of dye in solution (mg/L), Q_e is the amount of dye adsorbed per unit weight of adsorbent at equilibrium (mg/g), Q_m is the monolayer adsorption capacity (mg/g) and b is related with the energy of the adsorption (L/mg). A plot of $\log Q_e$ versus $\log C_e$ enables to determine the constant K_f and the adsorption intensity n . K_f represents the quantity of dye adsorbed onto adsorbent for an equilibrium concentration. When ($1/n > 1$), the change in adsorbed concentration is greater than the change in the solute concentration. The Freundlich model is suitable for use with heterogeneous surface but can describe the adsorption data over a restricted range only. It is often found that when the Freundlich equation is fitted to data at higher and intermediate concentrations, since the Freundlich equation does not approach Henry's Law of ideal dilute solutions. The Langmuir and Freundlich isotherm models gave straight lines, intercepts and slopes were used to determine the values of Langmuir and Freundlich parameters as given in Table 1.

Table 1 lists the maximum adsorption capacity Q_m values for BR12 adsorption onto ABM at different temperatures. From the results, it is clear that the value of adsorption efficiency Q_m and adsorption energy b of the ABM increase when increasing the temperature. From the values it is concluded that the maximum adsorption corresponds to a saturated monolayer of adsorbate molecules on adsorbent surface with

constant energy and no transmission of adsorbate occurs in the plane of the adsorbent surface. The observed b value shows that the adsorbent prefers to bind acidic ions and that speciation predominates on adsorbent characteristics, when ion exchange as the predominant mechanism takes place in the adsorption of BR12, it confirms the endothermic nature process involved in the system. The Freundlich constant, K_f , increases with increase in temperature. The value of n is greater than 1 indicating the adsorption of BR12 onto ABM is favorable.

Thermodynamic parameters such as change in free energy ΔG^0 (kJ/mol), enthalpy ΔH^0 (kJ/mol) and entropy ΔS^0 (kJ/K mol) were determined using the following equations:

$$K_0 = \frac{C_{\text{solid}}}{C_{\text{liquid}}} \quad (6)$$

$$\Delta G = -RT \ln K_0 \quad (7)$$

$$\log K_0 = \frac{\Delta S^0}{2.303R} - \frac{\Delta H^0}{2.303RT} \quad (8)$$

where K_0 is the equilibrium constant, C_{solid} is the solid phase concentration at equilibrium (mg/L), C_{liquid} is the liquid phase concentration at equilibrium (mg/L), T is the temperature expressed in Kelvin and R is the gas constant.

The ΔH^0 and ΔS^0 values obtained from the slope and intercept of Van't Hoff plots are presented in Table 2. The values of ΔH^0 are within the range of 1–93 kJ/mol indicating the physisorption. From these results it is clear that physisorption is much more favorable for the adsorption of BR12. The positive values of ΔH^0 show the endothermic nature of adsorption and it governs the possibility of physical adsorption (Cestari et al., 2008; Mall et al., 2006; Altunşik et al., 2010). Since in case of physical adsorption, while increasing the temperature of the system, the extent of BR12 adsorption increases, this rules out the possibility of chemisorption (Khattri and Singh, 2009; Mohan et al., 2002). The low value of ΔH^0 suggests that the BR12 is physisorbed onto ABM adsorbent. The negative values of ΔG^0 (Table 2) shows that the adsorption is highly favorable and spontaneous. The positive values of ΔS^0 (Table 2) shows the increased disorder and randomness at the solid solution interface of BR12 with ABM adsorbent that brings about some structural changes in the BR12 and the ABM. The enhancement of adsorption capacity of the ABM at higher temperatures was attributed to the enlargement of pore size and activation of the adsorbent surface (Arami et al., 2008; Garg et al., 2003).

The essential characteristics of the Langmuir isotherm can be expressed in terms of dimensionless constant separation factor R_L given by Boujaady et al. (2011), Deniz and Karaman (2011):

Table 1 Statistical parameters/constants of Langmuir and Freundlich isotherms results.

Temp. (K)	Langmuir isotherm results			Freundlich isotherm results		
	R^2	Q_m (mg/g)	b	R^2	K_f	n
303	0.9989	64.13	0.176	0.9983	7.41	2.12
313	0.9987	64.95	0.194	0.9989	8.65	2.24
323	0.9991	66.73	0.204	0.9961	9.23	2.31
333	0.9990	67.87	0.214	0.9979	9.71	2.42

Table 2 Equilibrium constant and thermodynamic parameters for the adsorption of BR12 onto ABM adsorbent.

[BR12] ₀ (mg/L)	20	40	60	80	100	
ΔG^0 (kJ/mol)	303 K	-3.474	-2.563	-1.811	-1.285	-1.218
	313 K	-3.919	-2.791	-2.124	-1.431	-1.248
	323 K	-4.395	-3.166	-2.518	-1.599	-1.304
	333 K	-4.838	-3.342	-2.960	-1.831	-1.387
ΔH^0 (kJ/mol)	—	15.76	13.46	7.76	6.97	6.05
ΔS^0 (kJ/K mol)	—	31.78	26.56	24.28	20.47	18.67

Table 3 Dimensionless separation factor R_L .

[BR12] (mg/L)	Temperature (K)			
	303	313	323	333
20	0.484	0.463	0.427	0.403
40	0.321	0.301	0.272	0.253
60	0.240	0.223	0.200	0.184
80	0.191	0.177	0.157	0.146
100	0.159	0.147	0.130	0.120

$$R_L = \frac{1}{1 + bC_0} \quad (9)$$

where b is the Langmuir constant and C_0 is the highest initial dye concentration (mg/L). R_L values indicate the type of Langmuir isotherm and to be (Boujaady et al., 2011; Deniz and Karaman, 2011): Irreversible ($R_L = 0$), Favorable ($0 < R_L < 1$), Linear ($R_L = 1$) and Unfavorable ($R_L > 1$). The R_L values depicted in Table 3 range between 0.120 and 0.484 for the range of temperatures studied indicate that the adsorption of BR12 onto ABM adsorbent is favorable.

4. Conclusion

The results indicated that ABM is a promising new low cost adsorbent for the removal of BR12 from aqueous solutions. The equilibrium data have been analyzed. The results showed that the BR12 followed Langmuir and Freundlich isotherm models. Thermodynamic studies indicated that the dye adsorption onto ABM was a spontaneous, endothermic and physical reaction in nature.

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